

Effects of Nitrogen Addition and Fire on Plant Nitrogen Use in a Temperate Steppe

Hai-Wei Wei^{1,2}, Xiao-Tao Lü^{1*}, Fu-Mei Lü³, Xing-Guo Han^{1,3}

1 State Key Laboratory of Forest and Soil Ecology, Institute of Applied Ecology, Chinese Academy of Sciences, Shenyang, China, **2** Graduate University of Chinese Academy of Sciences, Beijing, China, **3** State Key Laboratory of Vegetation and Environmental Change, Institute of Botany, Chinese Academy of Sciences, Beijing, China

Abstract

Plant nitrogen (N) use strategies have great implications for primary production and ecosystem nutrient cycling. Given the increasing atmospheric N deposition received by most of the terrestrial ecosystems, understanding the responses of plant N use would facilitate the projection of plant-mediated N cycling under global change scenarios. The effects of N deposition on plant N use would be affected by both natural and anthropogenic disturbances, such as prescribed fire in the grassland. We examined the effects of N addition ($5.25 \text{ g N m}^{-2} \text{ yr}^{-1}$) and prescribed fire (annual burning) on plant N concentrations and N use characters at both species and community levels in a temperate steppe of northern China. We found that N addition and fire independently affected soil N availability and plant N use traits. Nitrogen addition increased aboveground net primary productivity (ANPP), inorganic N, and N uptake, decreased N response efficiency (NRE), but did not affect biomass-weighted N concentrations at community level. Prescribed fire did not change the community level N concentrations, but largely decreased N uptake efficiency and NRE. At the species level, the effects of N addition and fire on plant N use were species-specific. The divergent responses of plant N use at community and species levels to N addition and fire highlight the importance of the hierarchical responses of plant N use at diverse biological organization levels to the alteration of soil N availability. This study will improve our understanding of the responses of plant-mediated N cycling to global change factors and ecosystem management strategies in the semiarid grasslands.

Citation: Wei H-W, Lü X-T, Lü F-M, Han X-G (2014) Effects of Nitrogen Addition and Fire on Plant Nitrogen Use in a Temperate Steppe. *PLoS ONE* 9(3): e90057. doi:10.1371/journal.pone.0090057

Editor: Fei-Hai Yu, Beijing Forestry University, China

Received: November 27, 2013; **Accepted:** January 27, 2014; **Published:** March 3, 2014

Copyright: © 2014 Wei et al. This is an open-access article distributed under the terms of the Creative Commons Attribution License, which permits unrestricted use, distribution, and reproduction in any medium, provided the original author and source are credited.

Funding: This research was supported by the National Natural Science Foundation of China (41273094, 41173086, and 31170433) and the State Key Laboratory of Forest and Soil Ecology (LFSE2013-13). The funders had no role in study design, data collection and analysis, decision to publish, or preparation of the manuscript.

Competing Interests: The authors have declared that no competing interests exist.

* E-mail: lvxiaotao@iae.ac.cn

Introduction

Nitrogen (N) availability is the dominant limiting factor for primary productivity in most terrestrial ecosystems [1–4]. Consequently, the responses of plants to the variation in soil N availability would have great implications for plant growth and species competition. It is well established that plant N economy plays an important role in affecting plant productivity and ecosystem N cycling [5,6]. Both plant N concentration and N response efficiency (NRE) are widely used to evaluate the responses of plant N economy to changes of soil N availability. Changes of plant N concentration have great implications for their N use efficiency (NUE), which is generally defined as the dry mass productivity per unit N taken up from the soil [7]. N response efficiency is the amount of biomass produced per unit of plant available N in the soil [8–10]. While plant N concentration is an indicator that integrates plant physiological and morphological responses to soil N availability, NRE is an index that integrates ecological and biogeochemical responses of plants. There is increasing empirical evidence that both plant N use indices should be simultaneously examined as they involve different processes linking the plant-soil system [11].

Both increasing atmospheric N deposition and global warming would enhance soil N availability in terrestrial ecosystems [12,13], which exerts positive effects on plant growth and primary

productivity [1,14]. It should be noted that, however, the relationship between increased N availability and plant N use characters may not be straightforward in natural ecosystems. For instance, the changes of NUE in response to N availability would be the balance between the responses of its two components, NPP and the amount of N taken up by plants. Similarly, the responses of NRE would be determined by the relative changes of NPP and soil N availability. A decrease of NRE in response to N addition means that N addition shows relatively stronger impacts on soil N availability than NPP. Previous studies showed that N addition generally resulted in an increase of plant N concentration and a decrease of NRE at species level [8,11,15,16]. It has been suggested that NRE would be a more representative measurement of plant responses to increasing soil N availability [11].

Most of previous studies mainly focused on changes of foliar N characters along nutrient availability gradients [6], which may not well mirror what is happening at higher biological organization levels. For example, there is significant variation of N concentrations between leaf and stem and also between aboveground and belowground parts at the individual plant level [17–18]. Furthermore, natural communities are always composed of plant species with various nutrient use strategies [19] and different nutrient limitation status [20]. There is increasing evidence that the response of plant N use to changes of nutrient availability would

vary significantly among diverse biological scales, from foliar to community level [11].

Environmental disturbances and ecosystem management strategies may also affect plant N use through their impacts on soil N availability [21–23] and community composition [24–26]. Moreover, they have the potential to alter the magnitude and direction of the impacts of N addition on plant N use [27]. Fire, as an anthropogenic or natural disturbance, exerts great control on community composition and ecosystem functioning in several terrestrial ecosystems, including grassland [25,28,29]. Reduction in fire frequency is considered as one of the major reasons for the increases of woody plant abundance in grasslands all over the world [26,30]. Prescribed burning is used as a strategy to prevent shrub encroachment in grasslands [31]. Results from a meta-analysis showed that short-term fire would enhance soil N availability while long-term repeated fire would lead to a decrease of soil N [21]. Similarly, the effects of fire on plant N status and use characters would also depend on its frequency. In the short-term, fire would increase plant NUE due to its stimulation on photosynthesis [32], whereas plant NUE would decrease under long-term repeated burning [22]. Given the interactive effects of fire and N addition on soil nutrient availability and plant production as observed in other studies [27], it is reasonable to expect that they may interact to affect plant N use. However, empirical evidence for their interactive effects on plant N use is still scarce.

The main objective of this study was to determine the effects of N addition, fire, and their interaction on plant N concentrations and N use characters at both species and community levels. We conducted a manipulative experiment in a temperate steppe in northern China, which is a typical vegetation type in the Eurasian continent. Fire was an important natural disturbance factor before the 1980s in northern China, with a frequency of once every 3–5 years [33]. In previous studies, we found that N addition and fire exert influence on plant nutrient resorption [34] and C:N:P stoichiometry [35,36]. We tested the following hypotheses: (1) N addition would enhance plant N concentrations, but reduce plant N uptake efficiency and NRE. (2) Fire would also increase plant N concentrations, but decrease plant N uptake efficiency and NRE, as the short-term fire (annual burning for two years) usually enhance soil N availability. Furthermore, we expected that the responses of plant N use at species level would be different from that at community level.

Materials and Methods

Site description

This study was conducted near the Inner Mongolia Grassland Ecosystem Research Station (IMGERS, 116°42'E, 43°38'N), which is located in the middle reach of the Xilin River, northern China. Long-term (1970–2007) meteorological data indicated that the mean annual precipitation was approximately 345 mm and mean annual temperature was 0.3°C. The length of growing season is about 150 days. The soil of this site is classified as Haplic Calcisols according to the Food and Agriculture Organization classification. Mean bulk density (top 10 cm) is 1.3 g cm⁻³ and pH is 7.4. The vegetation is dominated by *Leymus chinensis* (Trin.) Tzvel., *Stipa grandis* P. Smirn., *Cleistogenes squarrosa* (Trin.) Keng., *Caragana microphylla* Lam., *Carex korshinskyi* Kom., and *Potentilla bifurca* L.

Experimental design

This study was conducted as part of the GFE (Grassland Fire Experiment) experimental setup near IMGERS. All necessary

permits were gained from Institute of Botany, Chinese Academy of Sciences and local government before the beginning of this study. The design consists of nine blocks distributed across a grassland field, with each block containing a set of 10 m×10 m plots representing fully crossed treatments of fire frequency, N addition, and mowing frequency. Plots were separated by 1 m buffers. The experiment setup was established in 2005, with all burning, N addition, and mowing treatments were started since 2006. Burning was carried out in early or late April each year before the start of growing season. Nitrogen in the form of NH₄NO₃ was added in the rainy days of late June each year. Only the treatments of two burning frequencies (never burned vs annual burning) and two N addition levels (0 vs 5.25 g N m⁻² yr⁻¹) that without mowing in 9 blocks were used in this study, with 4 treatments ×9 replications = 36 plots in all.

Sample collection and chemical analysis

Aboveground net primary production (ANPP) was determined by clipping all plants in randomly located 1 m×1 m quadrats in each plot during the peak of growing season in August 2007. All plant samples were oven-dried at 70°C for 48 h, and then weighed. Samples of each species per plot were combined and ground to pass through a 40-mesh sieve using a mechanical mill (Retsch MM 400, Retsch GmbH & Co KG, Haan, Germany). Total N concentrations were determined colorimetrically by the Kjeldahl acid-digestion method with an AlpKem auto-analyzer (Kjektec System 1026 distilling unit, Sweden) after extraction with sulfuric acid.

Three soil cores (topsoil, 0–10 cm) were sampled in each of those quadrats following plant sampling with a metal tube (5 cm in diameter) and pooled to obtain one composite soil sample per quadrat. To measure soil water content, a soil subsample from each quadrat was weighed before and after being oven dried at 105°C for 48 h. Soil inorganic N availability was determined by extracting 10 g of fresh, root-free soil with 50 mL 2M KCl. The soil-extractant mixture was shaken for 1 h and then filtered (Whatman No. 1 filter paper). NH₄⁺-N was measured with the salicylate method while NO₃⁻-N was measured using the cadmium reduction method on a FIAstar 5000 Analyzer (Foss Tecator, Denmark). Total soil inorganic N was the sum of NH₄⁺-N and NO₃⁻-N.

Calculations and statistical analysis

The community level biomass-weighted N concentration was calculated as:

$$N \text{ concentration}_{community} = \frac{\sum N \text{ concentration}_i \times ANPP_i}{ANPP} \quad (1)$$

Where i is the i th species of n species in the community.

N uptake efficiency was calculated as the proportion of available N in the soil that was taken up by the plants [8]

$$N \text{ uptake efficiency} = \frac{N \text{ uptake}}{N \text{ available}} \quad (2)$$

NRE was calculated ANPP relative to the pool of available soil N [8–10]:

$$NRE = \frac{ANPP}{N \text{ available}} \quad (3)$$

The Kolmogorov-Smirnov test was used to test for data normality.

Two-way ANOVAs with fire and N addition as the main factors were performed to examine main and interactive effects on soil inorganic N, soil water content, ANPP, N concentrations, N uptake, N uptake efficiency, and NRE at community level. Two-way ANOVAs were used to examine the effects of N addition and fire on the plant N use characters of five dominant species at species level. All statistical analyses were performed using SPSS software (version 13.0 for windows, SPSS Inc., Chicago, IL, USA).

Results

Community level responses

Nitrogen addition and fire significantly enhanced soil inorganic N by 94.2% ($P<0.001$) and 25.8% ($P<0.05$), respectively (Fig. 1a, Table 1). Fire significantly reduced soil water content ($P<0.001$), whereas N addition did not affect soil water content (Fig. 1b, Table 1). Nitrogen addition significantly increased ANPP by 36.3% ($P<0.001$; Fig. 1c, Table 1), whereas fire reduced ANPP by 14.3% ($P<0.05$; Fig. 1c; Table 1).

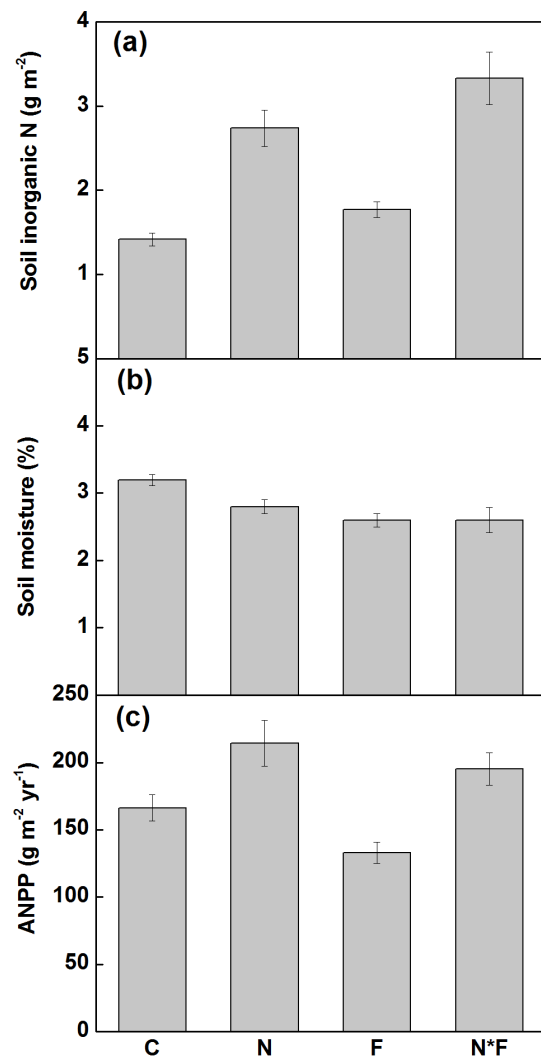


Figure 1. The effects of nitrogen addition and fire on (a) soil inorganic N concentrations, (b) soil water content, and (c) ANPP of plant community in a temperate steppe. C, control; N, nitrogen addition; F, fire; N*F, both N addition and fire. Data are adjusted means \pm 1SE, where $n=9$ for all treatments. doi:10.1371/journal.pone.0090057.g001

Both N addition and fire had no impacts on biomass-weighted N concentrations at community level (Fig. 2a; Table 1). N addition enhanced plant N uptake by 41.6% ($P<0.01$; Fig. 2b; Table 1), while fire had no effects on plant N uptake (Fig. 2b; Table 1). Both N addition and fire decreased N uptake efficiency by 22.9% ($P<0.01$) and 18.2% ($P<0.05$; Fig. 2c; Table 1). NRE was significantly reduced by both N addition (31.5%, $P<0.001$, Fig. 2c; Table 1) and fire (36.6%, $P<0.001$, Table 1). There was no interaction between N addition and fire for all variables (Table 1).

Species level responses

Nitrogen addition significantly enhanced the ANPP of *C. korshinskyi* ($P<0.01$) and *L. chinensis* ($P<0.05$; Table 1; Fig. 3), but showed no impacts on that of *A. cristatum*, *C. squarrosa*, and *S. grandis*. When the species were analyzed individually, plant N use varied greatly among five dominant species (Fig. 3). N addition had no impacts on N concentration for all species, while fire significantly increased N concentration of all species except for *C. squarrosa* (Table 1; Fig. 3). N addition and fire interacted to affect N concentration of *A. cristatum* ($P<0.05$) and *S. grandis* ($P<0.01$; Table 1; Fig. 3). While N addition significantly enhanced N uptake in two out of all the five species (*A. cristatum* and *C. korshinskyi*), fire significantly reduced N uptake of *C. korshinskyi* but did not affect that of the other species (Table 1; Fig. 3).

Nitrogen addition significantly decreased N uptake efficiency of all species ($P<0.05$) except for *C. korshinskyi* (Table 1, Fig. 3), while fire only decreased the N uptake efficiency of *C. korshinskyi* ($P<0.001$). N addition significantly decreased NRE for all species ($P<0.05$), whereas fire only decreased that of *A. cristatum* ($P<0.05$) and *C. korshinskyi* ($P<0.001$). For all the species, there was no interactive effect between N addition and fire on NRE (Table 1).

Discussion

Results from this study showed that N addition and fire directly affected soil N availability, ANPP, plant N uptake and use traits at both species and community levels. We found no interaction between N addition and fire in affecting plant N use. Generally, N addition increased ANPP and plant N uptake, decreased N uptake efficiency and NRE, with no significant effect on plant N concentrations. In contrast, fire showed positive effects on plant N concentrations (especially at species level), negative effects on NRE, and neutral effects on plant N uptake.

We found that both N addition and fire increased soil inorganic N concentrations. A meta-analysis showed that the soil inorganic N would increase immediately after fire and then gradually declined to the level before fire within one year [21]. All the burning plots had been burned for two years before soils were sampled in this study. Our results suggest that short-term fire would stimulate soil N turnover in the temperate steppe. While N is assumed to be the primary limiting nutrient in this ecosystem [14], we found negative effects of fire on ANPP even it stimulated soil N availability. We suspect that water availability would play an important role in mediating the effects of fire on ANPP in this semiarid grassland. Fire significantly reduced soil water content in this study. In the same experimental setup, Zhou et al. [33] also found that the soil moisture in unburned plots was significantly higher than that in the burned plots during the growing season. Furthermore, it has been shown that water addition instead of N addition significantly increased ANPP, indicating water availability would be the dominant limiting factor for plant growth in this ecosystem [37]. Consequently, it is reasonable to speculate that reduction of soil water availability after burning would account for

Table 1. Results (F-values) of two-way ANOVAs on the effects of nitrogen addition (N), fire (F), and their interaction on soil inorganic N, soil water content (SWC), ANPP, N concentration, N uptake, N uptake efficiency, and NRE at both community and species level in a temperate steppe.

Factor	Inorganic N	SWC	ANPP	N concentration	N uptake	N uptake efficiency	NRE
Community-level							
N	53.5***	2.5	24.2***	0	12.4**	7.9**	13.6***
F	6.0*	10.3**	6.2*	3.5	0.2	4.7*	20.7***
N*F	0.4	2.6	0.3	1.5	0.3	0.1	3
Species-level							
<i>Agropyron cristatum</i>							
N	-	-	3.6	0	4.6*	11.9**	10.9**
F	-	-	0.7	6.3*	0.1	1.5	5.8*
N*F	-	-	0.2	6.5*	0.5	0.2	2.5
<i>Carex korshinskyi</i>							
N	-	-	13.1**	3.5	13.5***	2.4	5.1*
F	-	-	16.3***	7.9**	8.4**	15.4***	27.6***
N*F	-	-	2.2	0.1	1.5	0.1	0.3
<i>Cleistogenes squarrosa</i>							
N	-	-	0.6	0.5	1	11.2**	12.5**
F	-	-	0.3	2.9	0	0.7	2.1
N*F	-	-	5.2*	2	5.4*	1.3	0.4
<i>Leymus chinensis</i>							
N	-	-	5.2*	2.3	0.9	7.0*	6.4*
F	-	-	0	6.7*	2.3	0.2	2.3
N*F	-	-	0.4	0	0	0	0
<i>Stipa grandis</i>							
N	-	-	0.1	1.9	0	5.1*	5.0*
F	-	-	3.8	18.2***	1.4	2.3	4.5
N*F	-	-	2	12.3**	3.7	0.7	0.1

The F-values are presented, together with their level of significance.

* $P < 0.05$;

** $P < 0.01$;

*** $P < 0.001$.

For all variables error $df = 36$.

doi:10.1371/journal.pone.0090057.t001

the negative effects of fire on community ANPP as observed in the present study.

Nitrogen addition would elevate soil inorganic N availability and thus increase ANPP, which had been generally tested and confirmed by previous studies [1,14,16]. Similarly, our results showed that N addition ($5.25 \text{ g N m}^{-2} \text{ yr}^{-1}$) significantly enhanced soil inorganic N by 94.2%, and increased ANPP at both species and community levels. In the same area, Bai et al. [14] reported that N addition, even as low as $1.75 \text{ g N m}^{-2} \text{ yr}^{-1}$, would significantly increase ANPP. It is notable that, however, the effects of N addition on ANPP would be largely dependent on soil water availability [37]. The positive effects of N addition on the growth of dominant species and the ANPP of whole community as observed in this study indicated that primary production was limited by N availability in this area. Given the projected increasing atmospheric N deposition, the primary productivity would be facilitated following the increasing N availability. We found no interactive effect of N addition and fire on primary production at community level, implying that the positive effects of N deposition on primary production would not be affected by

natural fire or prescribed burning in the temperate steppe of northern China.

Our results showed that N addition significantly increased plant N uptake by 41.6%, whereas fire had no effects on plant N uptake. The positive effects of N addition on plant N uptake were resulted from the enhancement of ANPP, as we found no impacts of N addition on plant N concentrations at both species and community levels. Previous studies showed that fire would increase N concentrations in plant tissues [27]. In this study, we found two-year annual fire significantly enhanced biomass-weighted plant N concentrations at community level, which is consistent with the results of our previous study which was carried out in the same ecosystem [35]. Due to the offset between the increases of plant N concentration and the decreases of ANPP following fire, annual burning showed no significant effects on plant N uptake. Fire showed negative influence on N uptake efficiency as it had no effects on plant N uptake but enhanced soil inorganic N. It has been reported that N addition would decrease N uptake efficiency, as the increases of N uptake cannot commensurate with the increases of soil inorganic N in a peatland and a managed

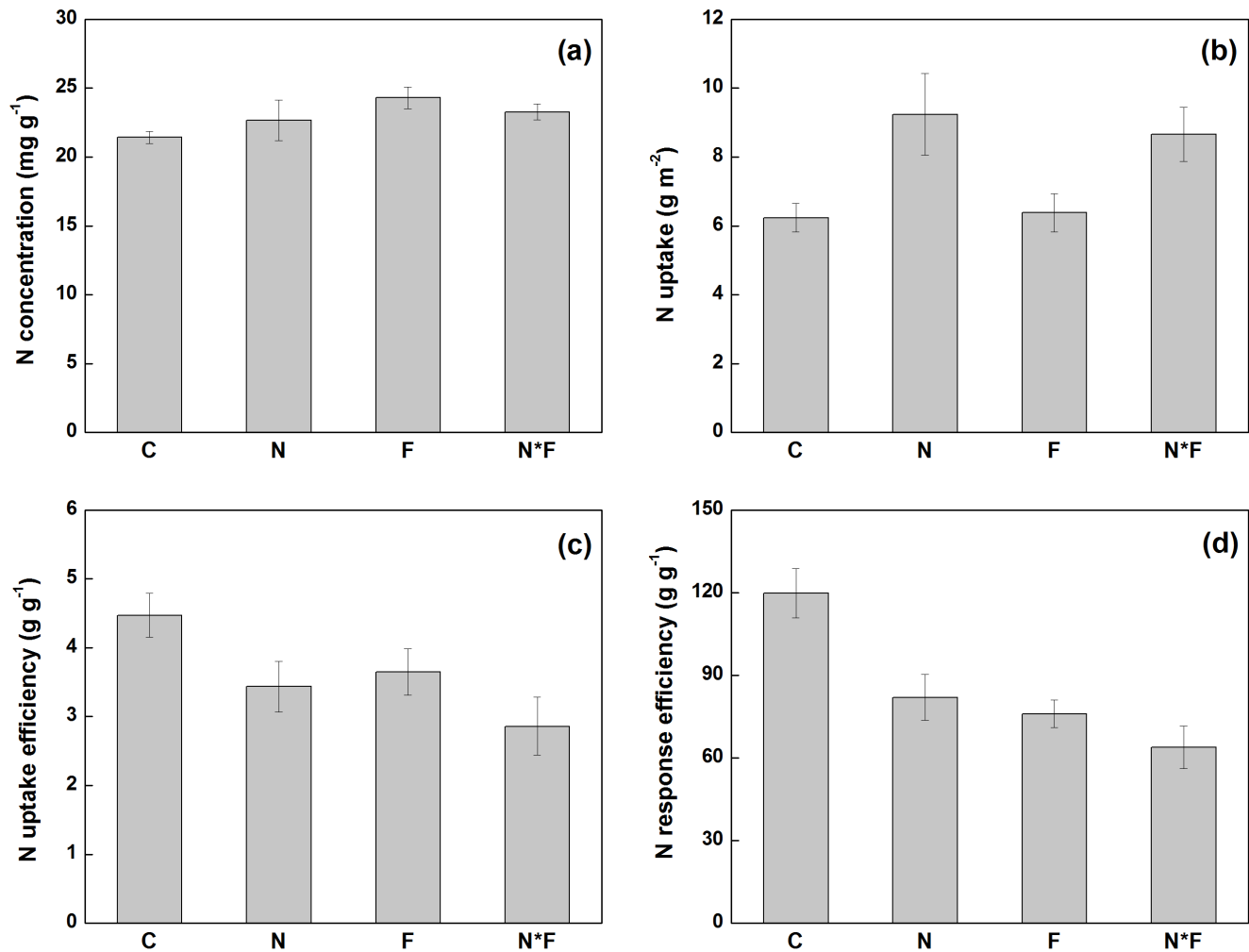


Figure 2. The effects of nitrogen addition and fire on (a) biomass-weighted N concentration, (b) N uptake, (c) N uptake efficiency, and (d) N response efficiency of plant community in a temperate steppe. C, control; N, nitrogen addition; F, fire; N*F, both N addition and fire. Data are adjusted means \pm 1SE, where $n=9$ for all treatments. doi:10.1371/journal.pone.0090057.g002

grassland [11,15]. Similarly, we found negative effects of N addition on plant N uptake efficiency at community level in this temperate steppe.

Previous studies reported that NUE often decreased with the increase of soil N availability [6,9,19], but not in other studies [38–40]. Contrary to our hypothesis, results from this study showed that N addition did not affect plant N concentrations at both species and community levels, implying that two-year N addition with the rate of $5.25 \text{ g N m}^{-2} \text{ yr}^{-1}$ did not affect NUE of plants in this temperate steppe. Our results were consistent with Reich et al. [41], who reported that increasing N availability did not change NUE in the forest ecosystem, but were different from Iversen et al. [11], who found that N addition ($6 \text{ g N m}^{-2} \text{ yr}^{-1}$) resulted in a decrease of NUE in peatlands. In temperate grassland, Yuan et al. [40] found that the responses of NUE to N addition were dependent on N addition rates. On the one hand, in our study, the N addition rate was $5.25 \text{ g N m}^{-2} \text{ yr}^{-1}$, which might be not enough to induce a decrease of NUE. On the other hand, our results indicate that the efficiency of temperate steppe to use N would be consistent in the face of increasing N deposition, at least at the rate of $5.25 \text{ g N m}^{-2} \text{ yr}^{-1}$. Our results that two years annual burning significantly enhanced biomass-weighted N con-

centrations at community level, indicating that fire would have negative effects of NUE in this community. The enhancement of plant N concentrations would have great means for litter quality and decomposition, as the litter decomposition rate and nutrient release are usually positive related to the N concentration in the litter [5,24,42].

Consistent with our hypothesis and findings from other studies [11,15], our results showed that N addition reduced NRE. Based on a natural soil N availability gradient in the semiarid grassland, Yuan et al. [43] found that NRE decreased monotonically with increasing N availability. The lower magnitude of the responses of ANPP to N addition than that of soil inorganic N availability resulted in a negative effect of N addition on NRE. Fire also led to a negative effect on plant NRE at community level, due to its negative effect on ANPP and positive effect on soil inorganic N. The negative effects of N addition and fire on community NRE suggests that part of soil inorganic N could not be taken up into plant biomass. Thus, more soil inorganic N would be lost by leaching and volatilization after N addition or burning. There is evidence that annually burning induce N loss through volatilization during fire [44], erosion, leaching, and denitrification [45]. It is notable that, the negative effects of N addition and fire on plant

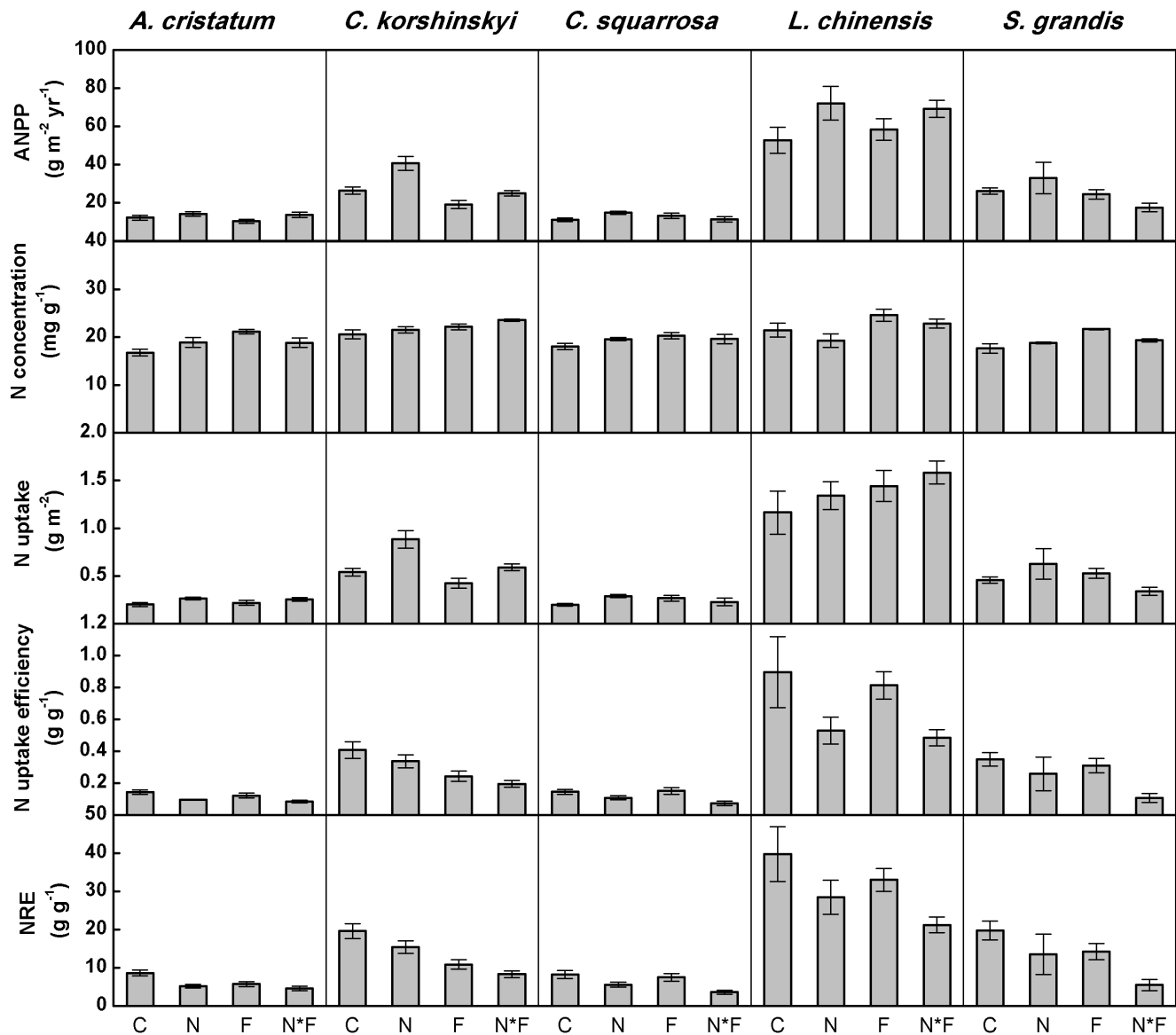


Figure 3. The effects of N addition and fire on plant N use of dominant species in a temperate steppe. C, control; N, nitrogen addition; F, fire; N*F, both N addition and fire. Data are adjusted means \pm 1SE, where $n=9$ for all treatments. doi:10.1371/journal.pone.0090057.g003

community NRE were independent, rather than additive. Together, our results indicate that the temperate steppe would be less efficient in absorbing the increased soil inorganic N into plant biomass after N addition and burning.

The dominant species examined in this study greatly varied in their N use traits. In the control plots, the NRE of *L. chinensis* was almost five times higher than that of *A. cristatum* and *C. squarrosa*. The higher NRE of *L. chinensis* would help explain its dominant status in this N-limited grassland ecosystem. More soil N would be transferred to plant biomass if this ecosystem was monodominated by *L. chinensis*, as indicated by its higher NRE status. Consequently, any changes of community composition following N addition and fire would influence plant N use at community level. Moreover, the responses of plant N economy to N addition and fire were species-specific. While N addition significantly enhanced the N uptake of *A. cristatum* and *C. korshinskyi*, it showed no impacts on that of other species. Similarly, fire reduced NRE of

A. cristatum and *C. korshinskyi*, but had no effects on other species. The idiosyncratic responses of different species to N addition and fire suggest that *A. cristatum* and *C. korshinskyi* were more sensitive to changes of soil N availability resulting from N addition and annual burning. Our results have great implications for the plant-mediated changes of biogeochemical cycling in response to global changes and ecosystem management strategies.

We hypothesized that the responses of plant N use traits to N addition and fire would be different at species and community levels. When all the dominant species were analyzed together, we found that the responses of plant N concentrations and NRE to N addition and fire did not differ ranging from species to community level. Nitrogen addition did not alter plant N concentrations at both species and community levels, whereas it did reduce plant NRE at both levels. Fire decreased plant N concentrations and NRE at community level, while its effects on species-level plant N concentrations and NRE varied among

different species. However, when those species were analyzed individually, we did find divergent responses of plant N use traits at species and community levels, which would be mainly resulted from the species-specific responses of plants to N addition and fire. In a recent study, Iversen et al. [11] reported that plant N use traits differed at different biological organizations ranging from individual plant tissue to the entire community in peatland ecosystems. Our study presented strong evidence for the hierarchical responses of plant N use to N addition and fire at different biological scales, especially for species and community levels in a grassland ecosystem.

Conclusions

Our results showed that both N addition and fire directly affected soil N availability and plant N use traits, with modest interaction between them. N addition increased ANPP, inorganic N, and N uptake, decreased NRE, and did not affect plant N concentrations. Prescribed fire, as an important management

strategy in grasslands, enhanced plant N concentrations and decreased N uptake efficiency and NRE. The responses of plant N use to both N addition and fire would be different at species and community levels, mainly due to the species-specific responses. The divergence of plant N use traits among different plant species indicates that any factors leading to changes of community composition would indirectly alter plant N use at community level.

Acknowledgments

We thank Wu Wei-Jun, Li Qiang, and Li Li for assistance with field and laboratory work, two anonymous reviewers for their helpful comments on the manuscript, and Matthew Simmons for language improvement.

Author Contributions

Conceived and designed the experiments: XH XL FL. Performed the experiments: FL XL HW. Analyzed the data: XL HW. Contributed reagents/materials/analysis tools: FL XH XL. Wrote the paper: HW XL.

References

1. LeBauer DS, Treseder KK (2008) Nitrogen limitation of net primary productivity in terrestrial ecosystems is globally distributed. *Ecology* 89: 371–379.
2. Perakis SS, Kellogg CH (2007) Imprint of oaks on nitrogen availability and delta N-15 in California grassland-savanna: a case of enhanced N inputs? *Plant Ecol* 191: 209–220.
3. Cole L, Buckland SM, Bardgett RD (2008) Influence of disturbance and nitrogen addition on plant and soil animal diversity in grassland. *Soil Biol Biochem* 40: 505–514.
4. Vitousek PM, Howarth RW (1991) Nitrogen limitation on land and in the sea: How can it occur? *Biogeochemistry* 13: 87–115.
5. Vitousek P (1982) Nutrient cycling and nutrient use efficiency. *Am Nat* 119: 553–572.
6. Aerts R, Chapin FS (2000) The mineral nutrition of wild plants revisited: A re-evaluation of processes and patterns. *Adv Ecol Res* 30: 1–47.
7. Hirose T (2011) Nitrogen use efficiency revisited. *Oecologia* 166: 863–867.
8. Pastor J, Bridgman SD (1999) Nutrient efficiency along nutrient availability gradients. *Oecologia* 118: 50–58.
9. Shaver GR, Melillo JM (1984) Nutrient budgets of marsh plants: efficiency concepts and relation to availability. *Ecology* 65: 1491–1510.
10. Bridgman SD, Pastor J, McLaugherty CA, Richardson CJ (1995) Nutrient-Use-Efficiency: A litterfall index, a model, and a test along a nutrient availability gradient in North-Carolina peatland. *Am Nat* 145: 1–21.
11. Iversen CM, Bridgman SD, Kellogg LE (2010) Scaling plant nitrogen use and uptake efficiencies in response to nutrient addition in peatlands. *Ecology* 91: 693–707.
12. Melillo JM, Stuedler PA, Aber JD, Newkirk K, Lux H, et al. (2002) Soil warming and carbon-cycle feedbacks to the climate system. *Science* 298: 2173–2176.
13. Wallenda T, Kottke I (1998) Nitrogen deposition and ectomycorrhizas. *New Phytol* 139: 169–187.
14. Bai Y, Wu J, Clark CM, Naeem S, Pan Q, et al. (2010) Tradeoffs and thresholds in the effects of nitrogen addition on biodiversity and ecosystem functioning: evidence from Inner Mongolia grasslands. *Glob Change Biol* 16: 358–372.
15. Keuter A, Hoefl I, Veldkamp E, Corre MD (2012) Nitrogen response efficiency of a managed and phytodiverse temperate grassland. *Plant Soil* 364: 193–206.
16. Chapin FS, Vitousek PM, Vancleve K (1986) The nature of nutrient limitation in plant-communities. *Am Nat* 127: 48–58.
17. Lu XT, Freschet GT, Flynn DFB, Han XG (2012) Plasticity in leaf and stem nutrient resorption proficiency potentially reinforces plant-soil feedbacks and microscale heterogeneity in a semi-arid grassland. *J Ecol* 100: 144–150.
18. Freschet GT, Cornelissen JHC, van Logtestijn RSP, Aerts R (2010) Substantial nutrient resorption from leaves, stems and roots in a subarctic flora: what is the link with other resource economics traits? *New Phytol* 186: 879–889.
19. Chapin FS (1980) The mineral nutrition of wild plants. *Ann Rev Ecol Syst* 11: 233–260.
20. Koerselman W, Meuleman AFM (1996) The vegetation N:P ratio: A new tool to detect the nature of nutrient limitation. *J Appl Ecol* 33: 1441–1450.
21. Wan SQ, Hui DF, Luo YQ (2001) Fire effects on nitrogen pools and dynamics in terrestrial ecosystems: A meta-analysis. *Ecol Appl* 11: 1349–1365.
22. Reich PB, Peterson DW, Wedin DA, Wragg K (2001) Fire and vegetation effects on productivity and nitrogen cycling across a forest-grassland continuum. *Ecology* 82: 1703–1719.
23. Pilkington MG, Caporn SJM, Carroll JA, Cresswell N, Phoenix GK, et al. (2007) Impacts of burning and increased nitrogen deposition on nitrogen pools and leaching in an upland moor. *J Ecol* 95: 1195–1207.
24. Watson PJ, Bradstock RA, Morris EC (2009) Fire frequency influences composition and structure of the shrub layer in an Australian subcoastal temperate grassy woodland. *Aust Ecol* 34: 218–232.
25. Bowman DM, Balch JK, Artaxo P, Bond WJ, Carlson JM, et al. (2009) Fire in the Earth system. *Science* 324: 481–484.
26. Bond WJ, Keeley JE (2005) Fire as a global 'herbivore': the ecology and evolution of flammable ecosystems. *Trends Ecol Evol* 20: 387–394.
27. Britton AJ, Helliwell RC, Fisher JM, Gibbs S (2008) Interactive effects of nitrogen deposition and fire on plant and soil chemistry in an alpine heathland. *Environ Poll* 156: 409–416.
28. Keeley JE, Pausas JG, Rundel PW, Bond WJ, Bradstock RA (2011) Fire as an evolutionary pressure shaping plant traits. *Trends Plant Sci* 16: 406–411.
29. Bradshaw SD, Dixon KW, Hopper SD, Lambers H, Turner SR (2011) Response to Keeley et al.: Fire as an evolutionary pressure shaping plant traits. *Trends Plant Sci* 16: 405–405.
30. Heisler JL, Briggs JM, Knapp AK, Blair JM, Seery A (2004) Direct and indirect effects of fire on shrub density and aboveground productivity in a mesic grassland. *Ecology* 85: 2245–2257.
31. Bowles ML, Jones MD (2013) Repeated burning of eastern tallgrass prairie increases richness and diversity, stabilizing late successional vegetation. *Ecol Appl* 23: 464–478.
32. Ojima DS, Schimel DS, Parton WJ, Owensby CE (1994) Long-term and short-term effects of fire on nitrogen cycling in tallgrass prairie. *Biogeochemistry* 24: 67–84.
33. Zhou L, Huang J, Lü F, Han X (2009) Effects of prescribed burning and seasonal and interannual climate variation on nitrogen mineralization in a typical steppe in Inner Mongolia. *Soil Biol Biochem* 41: 796–803.
34. Lü XT, Cui Q, Wang QB, Han XG (2011) Nutrient resorption response to fire and nitrogen addition in a semi-arid grassland. *Ecol Engin* 37: 534–538.
35. Cui Q, Lü XT, Wang QB, Han XG (2010) Nitrogen fertilization and fire act independently on foliar stoichiometry in a temperate steppe. *Plant Soil* 334: 209–219.
36. Lü XT, Lu FM, Zhou LS, Han X, Han XG (2012) Stoichiometric response of dominant grasses to fire and mowing in a semi-arid grassland. *J Arid Environ* 78: 154–160.
37. Lü XT, Kong DL, Pan QM, Simmons ME, Han XG (2012) Nitrogen and water availability interact to affect leaf stoichiometry in a semi-arid grassland. *Oecologia* 168: 301–310.
38. Aerts R, Decaluwe H (1994) Nitrogen use efficiency of *Carex* species in relation to nitrogen supply. *Ecology* 75: 2362–2372.
39. Yasumura Y, Hikosaka K, Matsui K, Hirose T (2002) Leaf-level nitrogen-use efficiency of canopy and understorey species in a beech forest. *Fun Ecol* 16: 826–834.
40. Yuan ZY, Li LH, Huang JH, Han XG, Wan SQ (2005) Effect of nitrogen supply on the nitrogen use efficiency of an annual herb, *Helianthus annuus* L. *J Integrat Plant Biol* 47: 539–548.
41. Reich PB, Grigal DF, Aber JD, Gower ST (1997) Nitrogen mineralization and productivity in 50 hardwood and conifer stands on diverse soils. *Ecology* 78: 335–347.
42. Blair JM, Parmelee RW, Beare MH (1990) Decay-rates, nitrogen fluxes, and decomposer communities of single-species and mixed-species foliar litter. *Ecology* 71: 1976–1985.
43. Yuan ZY, Li LH, Han XG, Chen SP, Wang ZW, et al. (2006) Nitrogen response efficiency increased monotonically with decreasing soil resource availability: a case study from a semiarid grassland in northern China. *Oecologia* 148: 564–572.

44. Olear HA, Seastedt TR, Briggs JM, Blair JM, Ramundo RA (1996) Fire and topographic effects on decomposition rates and N dynamics of buried wood in tallgrass prairie. *Soil Biol Biochem* 28: 323–329.
45. Neary DG, Klopatek CC, DeBano LF, Ffolliott PF (1999) Fire effects on belowground sustainability: a review and synthesis. *For Ecol Manag* 122: 51–71.